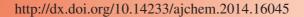
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Determination of Natural Radioactivity Levels of Some Concretes and Mineral Admixtures in Turkey

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Nine concrete samples with or without mineral admixtures have been analyzed for their naturally occurring radionuclide of 226 Ra, 232 Th and 40 K using HPGe gamma spectrometry. Also, concrete raw materials such as aggregate, cement, fly ash and blast furnace slag have been measured. The radioactivity values of 226 Ra, 232 Th and 40 K in fly ash and those of 226 Ra in blast furnace slag are higher than the corresponding world mean specific activities values which are 50, 50 and 500 Bq kg $^{-1}$ for 226 Ra, 232 Th and 40 K, respectively. The concretes activity concentrations have been found to change with supplementary cementitious materials of the concrete samples. This may be attributed to containing different proportions of different mineral additives such as fly ash and blast furnace slag which were found to contain high natural radionuclide concentrations according to the rest of raw material samples. In addition, radium equivalent activities (Ra_{eq}), γ -index (I $_{\gamma}$), α -index (I $_{\alpha}$), absorbed dose rate in air (D) and annual effective dose associated with the natural radionuclide were calculated to assess the radiation hazard of the natural radioactivity in the concrete mixture samples. The concentration of the natural radionuclides and the radium equivalent activity obtained in the present study are compared with the previous results. From these results, it can be seen that these building materials do not pose significant radiation hazards.

Keywords: Natural radioactivity, Concrete, Fly ash, Blast furnace, Cement.

INTRODUCTION

In its most basic form, concrete is a mixture of portland cement, sand, coarse aggregates and water. The principle cementitious material in concrete is portland cement. Today, most concrete mixtures contain supplementary cementitious materials that make up a portion of the cementitious component in concrete. Supplementary cementitious materials such as fly ash and granule blast furnace slag enable the concrete industry to use hundreds of millions of tons of by product materials that would otherwise be used as land-filled as in the west¹. In recent years fly ash and granule blast furnace have been used as a replacement of sand and cement in premixed concrete, manufacture of blended fly ash portland cement, aerated concrete, fly as clay bricks and blocks and for the filling of underground cavities etc.². The building industry requires large quantities of low-cost materials and new products that may be substitutes for the widely used natural products as conventional building materials. However, such new building materials may contain significant quantities of naturally or technologically enhanced levels of radioactivity³. The presence of these radioisotopes in materials causes external exposure to people who live in the building⁴.

Natural radioactivity in concrete materials comes mainly from uranium (²³⁸U) and thorium (²³⁸Th) series and the radioactive isotopes of potassium (⁴⁰K). In the uranium series, the decay chain segment starting from radium (²²⁶Ra) is radiological the most important and, therefore, reference is often made to radium instead of uranium. The knowledge of natural radioactivity in these materials is then important for determining the amount of public exposure because people spend most of their time (about 80 %) indoors⁵. This knowledge is important in the assessment of possible radiological hazards to human health and essential for the development of standards and guidelines in using and these materials.

In the recent years, there has been an increasing interest in the study of radioactivity in various building and their raw materials⁶⁻¹².

The aim of this paper is to measure the natural activity concentrations of some concrete samples. The natural activity concentrations in two different types of concrete (MF and MB) and their initial raw materials (fly ash and granule blast furnace slag) were also measured in order to assess each material's contribution to total concrete activity. The results obtained are compared with the results available in some other countries of the world.

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EXPERIMENTAL

The composition of the mortars all derive from the composition of standardized mortar (concrete) as specified in the TS EN 196-1 standard. All concrete is used in normalized sand CEN EN 196-1 and the ordinary Portland cement, CEM I 42.5 R. This cement was mixed with two mineral additives; fly ash and granule blast furnace slag in some mortars. Concrete samples were prepared by Hobart mixer according to TS EN 196-1. Mortar was placed on the jolting Table is blocked in a mould to be $4 \text{ cm} \times 4 \text{ cm} \times 16 \text{ cm}$. After 24 h of curing 20 °C by 90 % relative humidity, the samples were removed from the molds and placed in standard curing condition in statured lime water at 20 ± 1 °C. Three main different concrete types were prepared. For the basic type, without mineral additives, "ordinary" concrete mixture (MO, 0 %) consists of 450 g cement, 1350 aggregate and 225 g water (proportion 1:3:0.5 respectively). The other two types were produced by adding supplementary cementitious materials to the cement. In case of the fly ash and granule blast furnace slag concretes, respectively, (MF) and (MB) were incorporated in the cementitious materials at proportions of 5, 10, 20 and 30 % by weight. Finally, a total of nine concrete types were obtained. Again, all concrete mixtures have the same weight for total cementitious materials (cement and mineral additives) which equals to 450 g. The compositions of the concretes were listed in Table-1².

Radioactivity measurements: Each sample was homogenized and dried in a temperature controlled furnace at 105 °C for 24 h to remove moisture and sieved through a 2 mm mesh. About 120 g of samples were sealed in gas tight, radon impermeable, cylindrical polyethylene plastic containers (5.5 cm diameter and 5 cm height) for γ -activity analysis. Before measurements, the containers were kept sealed for 4 weeks in order to reach the equilibrium of 226 Ra and its short lived progeny.

Gamma spectrometry measurements were conducted with a coaxial HPGe detector of 55 % relative efficiency and a resolution of 1.9 keV at the 1332 keV gamma of ⁶⁰Co (Ortec, GEM55P4-95 model). The detector was shielded in a 10 cm thick lead well internally lined with 2 mm Cu foils. The spectrum analysis was performed using computer software Genie 2000 obtained from Ortec. A performance test using the certified reference samples (IAEA-375, IAEA, Vienna) of known activities was conducted for checking the efficiency calibration of the system. The specific activities of these samples were in accordance with their certified values within errors not exceeding 10 %. The quality assurance of the measurements was carried out by periodical efficiency and energy calibration and repeated sample measurement. The counting time for each sample was selected to be 50,000 to obtain the gamma-spectrum with good statistics. To determine the background distribution in the environment around the detector, an empty container was counted in the same manner and in the same geometry as the samples. The background spectra were used to correct the net peak area of γ-rays of the measured

The γ -ray transitions of energies 351.9 keV (214 Pb) and 609.3 keV (214 Bi) were used to determine the activity concentration of the 226 Ra series. The γ -ray lines at 911.1 keV (228 Ac)

and 583.1 keV (208 Tl) were used to determine the activity concentration of the 232 Th series. The activity concentrations of 40 K were measured directly through the gamma line emission at 1460.8 keV.

The activity concentrations for the natural radionuclides in the measured samples are computed using the following relation:

$$C = \frac{N}{\epsilon PMt} \left(Bq \ kg^{-1} \right) \tag{1}$$

where N is the net counting rate of the gamma ray, ε is the photo peak efficiency of the used detector, P is the absolute transition of gamma decay, t is the counting time in seconds and M the weight of the dried sample in kilogram.

Radium equivalent activity (Ra_{eq}): The distribution of ^{226}Ra , ^{232}Th and ^{40}K in the concrete materials was not uniform. Uniformity with respect to exposure to radiation was defined in terms of radium equivalent activity (Ra_{eq}) in Bq kg⁻¹ to compare the specific activity of materials containing different amounts of ^{226}Ra , ^{232}Th and ^{40}K . The radium equivalent activity is calculated through the following relation 13 :

$$Ra_{eq} = C_{Ra} + 1.43 C_{Th} + 0.077 C_{K}$$
 (2)

where C_{Ra} , C_{Th} and C_K are the activity concentrations of 226 Ra, 232 Th and 40 K in Bq kg⁻¹, respectively.

γ-Index (I_{γ}): Number of indices dealing with the assessment of the excess γ-radiation arising from building materials such as external and internal hazard indices and γ-concentration indices has been proposed by several investigators¹³⁻¹⁶. In this study, the γ-index (I_{γ}) is calculated as proposed by the European commission¹⁴.

$$I_{\gamma} = \frac{C_{Ra}}{300} + \frac{C_{Th}}{200} + \frac{C_{K}}{3000} \tag{3}$$

where C_{Ra} , C_{Th} and C_K are the activity concentrations of ^{226}Ra , ^{232}Th and ^{40}K in Bq kg $^{-1}$, respectively.

 α -Index (I_{α}): So far, several α -indices have been proposed to assess the exposure level due to radon inhalation originating from building materials ¹⁴. The α -index is determined by the following formula:

$$I_{\alpha} = \frac{C_{Ra}}{200 \, (\text{Bg kg}^{-1})} \tag{4}$$

where C_{Ra} (Bq kg⁻¹) is the activity concentration of ²²⁶Ra assumed in equilibrium with ²³⁸U. The recommended exemption and upper level of ²²⁶Ra activity concentrations in building materials are 100 and 200 Bq kg⁻¹, respectively, as suggested by ICRP¹⁷. These considerations are reflected in the α -index. The recommended upper limit concentration of ²²⁶Ra is 200 Bq kg⁻¹, for which $I_{\alpha}=1$.

Calculation of air-absorbed dose rates: There is concern that some of the buildings will cause excessive radiation doses to the total body due to γ -rays emitted by 214 Pb and 214 Bi progeny of 226 Ra and 232 Th decay chains and 40 K also contributes to the total body radiation dose. The absorbed dose rate in indoor air due to γ -ray emission from activity concentrations of 226 Ra, 232 Th and 40 K is estimated using the following formula provided by UNSCEAR 18 and EC 14 .

$$D = \alpha C_{Ra} + \beta C_{Th} + \gamma C_K$$
 (5)

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where α , β and γ are the dose rates per unit activity concentrations of Ra, Th and K (nGy h⁻¹/Bq kg⁻¹); C_{Ra} , C_{Th} and C_{K} are the activity concentrations of Ra, Th and K (Bq kg⁻¹), respectively. The values of α , β and γ were taken to be 0.92, 1.1 and 0.08, respectively¹⁴.

Calculation of the annual effective dose: To estimate the annual effective dose, it must be taken into account: (a) the conversion coefficient (0.7 Sv Gy^{-1}) from absorbed dose in air to effective dose and (b) the indoor occupancy factor (about 80 % for Turkish population). Therefore, the effective dose rate in mSv y^{-1} unit is estimated by the following formula¹⁹:

AED (mSv y⁻¹) = D (nGy h⁻¹) × 8760 h y⁻¹
×
$$0.7 \times 0.8 \times 10^{-6}$$
 (6)

RESULTS AND DISCUSSION

The composition (in gram) of the nine concrete mixtures is summarized in Table-1. In Turkey, these types of concretes are commonly used as building material. The results of activity concentrations are given in Table-2 for the natural radio-nuclides of ²²⁶Ra, ²³²Th and ⁴⁰K in ordinary concrete MO and concretes loaded with different proportions of different mineral additives (MF and MB). It can also be seen from Table-2 that

the results increase with increasing percentage of additions for concrete mixture samples. When the concrete mixture samples were compared in their own right, it was observed that the ²²⁶Ra, ²³²Th and ⁴⁰K activity concentrations of concrete mixture samples increase with the addition of fly ash and blast furnace slag at percentages of 0, 5, 10, 20 and 30 % as shown Fig. 1. The specific activities values obtained in this study were found to be lower than recommended maximum level for all mixture concrete samples²⁰.

TABLE-1						
COMPOSITION (g) OF THE NINE CONCRETE MIXTURES						
				Blast		
Mortars	(%)	Cement	Water	Fly ash	furnace	Aggregate
					slag	
MO	0	450	225	0	0	1350
MF5	5	427.5	225	22.5	0	1350
MF10	10	405	225	45	0	1350
MF20	20	360	225	90	0	1350
MF30	30	315	225	135	0	1350
MB5	5	427.5	225	0	22.5	1350
MB10	10	405	225	0	45	1350
MB20	20	360	225	0	90	1350
MB30	30	315	225	0	135	1350

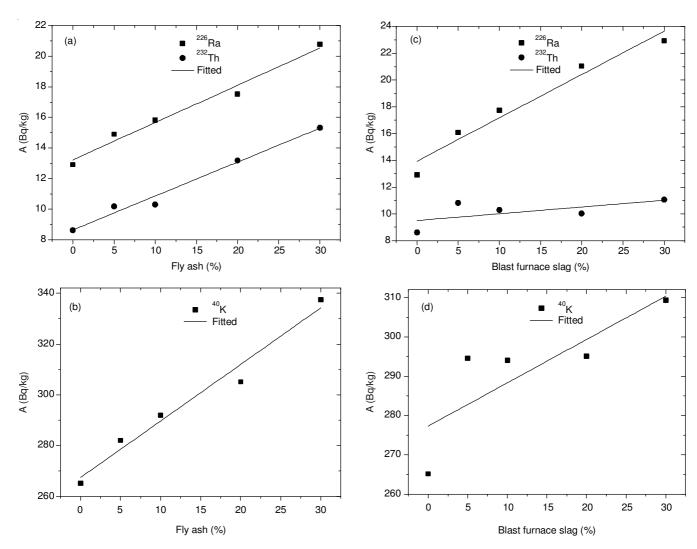


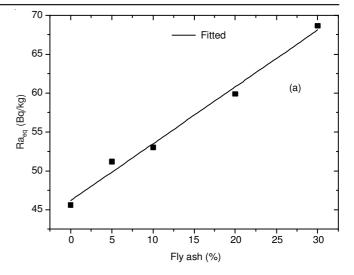
Fig. 1. Activity concentrations in concrete mixture samples versus per cent of values of fly ash (a, b) and blast furnace slag (c, d)

TABLE-2 ACTIVITY CONCENTRATIONS (Bq kg⁻¹) OF ²³²Th, ²²⁶Ra AND ⁴⁰K IN CONCRETE SAMPLES ACCORDING TO THE DIFFERENT PERCENTAGES AND CONCRETE RAW MATERIALS

Samples	Radionuclide concentrations (Bq/kg)				
Samples	²²⁶ Ra ²³² Th		⁴⁰ K		
MO	12.91 ± 0.59	8.61 ± 0.73	265.18 ± 7.02		
MF5	14.90 ± 0.65	10.18 ± 0.82	282.06 ± 7.49		
MF10	15.81 ± 0.67	10.30 ± 0.83	291.97 ± 7.68		
MF20	17.53 ± 0.70	13.19 ± 0.88	305.16 ± 7.83		
MF30	20.78 ± 0.79	15.32 ±1.00	337.45 ± 8.71		
MB5	16.08 ± 0.73	10.81 ± 0.89	294.55 ± 8.20		
MB10	17.74 ± 0.69	10.29 ± 0.88	294.02 ± 7.92		
MB20	21.06 ± 0.75	10.03 ± 0.79	295.05 ± 7.29		
MB30	22.95 ± 0.74	11.06 ± 0.87	309.32 ± 8.10		
Aggregate	9.51 ± 0.49	8.00 ± 0.65	322.91 ± 7.37		
Fly ash	114.67 ± 2.08	87.86 ± 2.71	898.29 ± 16.41		
Blast furnace slag	186.69 ± 2.38	35.87 ± 1.67	295.91 ± 9.08		
Cement	26.09 ± 0.83	8.40 ± 0.83	168.42 ± 5.76		

In addition, the natural radionuclide concentrations of concretes 'raw materials such as cement, fly ash, blast furnace slag and aggregate were listed in Table-2. It is observed that the measured activity concentration values of ²²⁶Ra, ²³²Th and ⁴⁰K in fly ash, blast furnace slag and ⁴⁰K in aggregate are higher according to the other raw samples. As can be seen from Table-2, the radioactivity values of ²²⁶Ra, ²³²Th and ⁴⁰K in fly ash and those of ²²⁶Ra in blast furnace slag are higher than the corresponding world mean specific activities values which are 50, 50 and 500 Bq kg⁻¹ for ²²⁶Ra, ²³²Th and ⁴⁰K, respectively²⁰. Fly ash and blast furnace seems to be problematic from radionuclides content point of view. However, fly ash is used in Turkish cement industry as an addition to Portland-fly ash cement in an amount not exceeding 35 % (6-35 %)²¹. As a result, the natural activities values of the above-mentioned radionuclides in ordinary concrete (MO), concretes loaded with different proportions of different mineral additives (MF and MB), cement and aggregate are lower than the world mean values. From the point of view radionuclide, the mixing of fly ash and blast furnace slag with concrete components prepared in this work is suitable for the walls of the new buildings.

The mean calculated Ra_{eq} values are shown in Table-3 for the MO, MF and MB and for building raw materials (fly ash, blast furnace, cement and aggregate). The mean Ra_{eq} values varies from 45.64 ± 2.40 to 68.67 ± 2.89 Bq kg⁻¹, 45.64 ± 2.63



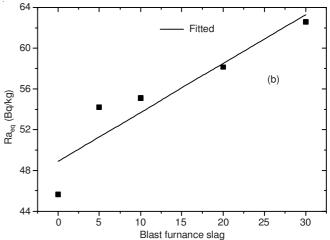


Fig. 2. Radium equivalent (Ra_{eq}) activities in concrete mixture samples versus per cent of values of fly ash (a) and blast furnace slag (b)

to 62.58 ± 2.61 Bq kg⁻¹ for MF and MB, respectively. It can be seen that the highest mean values of Ra_{eq} activity were determined in fly ash (309.48 \pm 7.22 Bq kg⁻¹) and blast furnace (260.77 \pm 5.47 Bq kg⁻¹). However, the obtained results indicate that the mean Ra_{eq} values in this work are less than the recommended maximum levels for building materials to be used for house, which is < 370 Bq kg⁻¹ ^{18,19,21,22}. The results were also shown graphically in Fig. 2 for concrete mixture samples

TABLE 3
RADIUM EQUIVALENT ACTIVITY, γ-INDICES, α-INDICES, THE DOSE RATE AND ANNUAL EFFECTIVE DOSE FOR CONCRETE SAMPLES ACCORDING TO THE DIFFERENT PERCENTAGES AND CONCRETE RAW MATERIALS

Samples	Ra _{eq} (Bq/kg)	${ m I}_{\gamma}$	${ m I}_{lpha}$	D (nGy h ⁻¹)	AED (mSv y ⁻¹)
MO	45.64 ± 2.17	0.17	0.06	42.56 ± 1.91	0.21
MF5	51.18 ± 2.40	0.19	0.07	47.47 ± 2.10	0.23
MF10	53.02 ± 2.45	0.20	0.08	49.23 ± 2.14	0.24
MF20	59.89 ± 2.56	0.23	0.09	55.05 ± 2.24	0.27
MF30	68.67 ± 2.89	0.26	0.10	62.97 ± 2.52	0.31
MB5	54.22 ± 2.63	0.21	0.08	50.25 ± 2.31	0.25
MB10	55.09 ± 2.56	0.21	0.09	51.16 ± 2.24	0.25
MB20	58.12 ± 2.44	0.22	0.11	54.01 ± 2.14	0.26
MB30	62.58 ± 2.61	0.23	0.11	58.03 ± 2.29	0.28
Aggregate	45.81 ± 1.99	0.18	0.05	43.38 ± 1.76	0.21
Fly ash	309.48 ± 7.22	1.12	0.57	274.01 ± 6.21	1.34
Blast furnace slag	260.77 ± 5.47	0.90	0.93	234.88 ± 4.75	1.15
Cement	51.07 ± 2.46	0.19	0.13	46.72 ± 2.14	0.23

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TABLE-4 COMPARISON OF ACTIVITY CONCENTRATIONS (Bq kg⁻¹) AND THE RADIUM EQUIVALENT (Ra_{eq}) ACTIVITIES WITH PUBLISHED RESULTS IN CEMENT (A), FLY ASH (B), BLAST FURNACE SLAG (C), AGGREGATE (D) AND CONCRETE (E) SAMPLES FROM DIFFERENT COUNTRIES

Column	Countries	Radio	nuclide concentrations (Bo		- Ra _{eq} (Bq kg ⁻¹)	References
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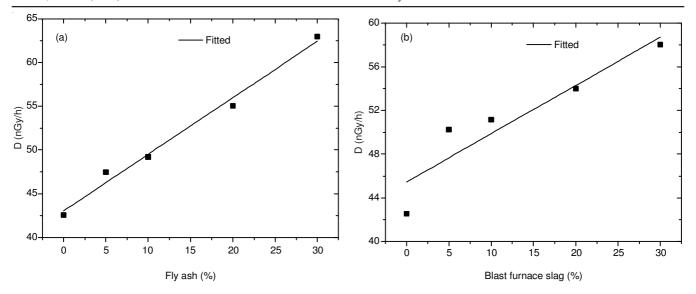


Fig. 3. Dose rate (D) values in concrete mixture samples versus per cent of values of fly ash (a) and blast furnace slag (b)

versus per cent of values of fly ash and blast furnace slag, respectively. Fig. 2 demonstrate same noticed feature in previous figure, such as increasing of radium equivalents with mineral additives. Table-4a-e lists the comparison between activity concentrations and radium equivalents (Bq kg⁻¹) in cement, fly ash, blast furnace, aggregate and concrete in different areas of the world, respectively. The comparison shows a reasonable agreement for cement, blast furnace slag, aggre-gate and mixture concretes samples in Table-4a, c, d, e, respectively. As shown in Table-4b, if we compare our results with those of values in different countries for fly ash, it is clear that ²²⁶Ra, ²³²Th and Ra_{eq} is good agreement, while ⁴⁰K is high in the present work.

The mean values of γ -index (I_{γ}) calculated from the measured activity concentration of ^{226}Ra , ^{232}Th and ^{40}K are listed in Table-3 (column 2) for all the studied samples. It is observed that the mean γ -index I_{γ} values do not exceed critical value of unity for all samples. These values of I_{γ} are < 2, which corresponds to a dose criterion of 0.3 mSv y^{-1} . It is observed that mean I_{γ} of values for fly ash (1.12) and blast furnace slag (0.90) samples are higher than the values of the rest of the building raw materials and, the concrete mixture samples.

The α -index (I_{α}) was calculated from eqn. (4) and listed in Table-3 (column 3). As shown in Table-3, the calculated values of I_{α} for the studied samples are less than unity.

The estimated absorbed dose rate (D) values for the studied samples are also shown in Table-3 (column 4). All the samples measured, except for fly ash (274.01 ± 6.21) and blast furnace slag (234.88 ± 4.75) , showed values below the recommended upper level, which is 84 nGy h⁻¹.

The obtained annual effective dose rate results are in Table-3 (column 5) and displayed in Fig. 3. The annual effective dose values for all samples except for fly ash $(1.34 \text{ mSv y}^{-1})$ and blast furnace slag $(1.15 \text{ mSv y}^{-1})$ are lower than the limit value, which is 0.46 mSv y^{-1} ¹⁸.

Conclusion

The specific activity of ²²⁶Ra, ²³²Th and ⁴⁰K have been measured in ordinary concrete MO, concretes loaded with

different proportions of different mineral additives (MF and MB) and building raw materials used in Turkey, using HPGe gamma spectrometry. The radioactivity values of ^{226}Ra , ^{232}Th and ^{40}K in fly ash and those of ^{226}Ra in blast furnace slag are higher than the corresponding world mean specific activities values, while the rest of the samples are less than the critical value of unity. In addition, the mean radium equivalent activity (Raeq), γ -index (I γ), α -index (I α), the indoor absorbed dose rate (D) and annual effective dose rate have been calculated for all samples. The observed these values except for annual effective dose values of fly ash and blast furnace slag are lower than the acceptable level. According to these results, mixture concrete samples examined in this study do not pose any significant source of radiation hazard and thus these materials could be safely used in building constructions.

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